Respiratory Effects of Exposure to Diesel Traffic in Persons with Asthma

James McCreanor, M.R.C.P., Paul Cullinan, M.D., Mark J. Nieuwenhuijsen, Ph.D.,
James Stewart-Evans, M.Sc., Eleni Malliarou, M.Sc., Lars Jarup, Ph.D.,
Robert Harrington, M.S., Magnus Svartengren, M.D., In-Kyu Han, M.P.H.,
Pamela Ohman-Strickland, Ph.D., Kian Fan Chung, M.D.,
and Junfeng Zhang, Ph.D.

ABSTRACT

BACKGROUND
Air pollution from road traffic is a serious health hazard, and people with preexisting respiratory disease may be at increased risk. We investigated the effects of short-term exposure to diesel traffic in people with asthma in an urban, roadside environment.

METHODS
We recruited 60 adults with either mild or moderate asthma to participate in a randomized, crossover study. Each participant walked for 2 hours along a London street (Oxford Street) and, on a separate occasion, through a nearby park (Hyde Park). We performed detailed real-time exposure, physiological, and immunologic measurements.

RESULTS
Participants had significantly higher exposures to fine particles (<2.5 μm in aerodynamic diameter), ultrafine particles, elemental carbon, and nitrogen dioxide on Oxford Street than in Hyde Park. Walking for 2 hours on Oxford Street induced asymptomatic but consistent reductions in the forced expiratory volume in 1 second (FEV₁) (up to 6.1%) and forced vital capacity (FVC) (up to 5.4%) that were significantly larger than the reductions in FEV₁ and FVC after exposure in Hyde Park (P = 0.04 and P = 0.01, respectively, for the overall effect of exposure, and P < 0.005 at some time points). The effects were greater in subjects with moderate asthma than in those with mild asthma. These changes were accompanied by increases in biomarkers of neutrophilic inflammation (sputum myeloperoxidase, 4.24 ng per milliliter after exposure in Hyde Park vs. 24.5 ng per milliliter after exposure on Oxford Street; P = 0.05) and airway acidification (maximum decrease in pH, 0.04% after exposure in Hyde Park and 1.9% after exposure on Oxford Street; P = 0.003). The changes were associated most consistently with exposures to ultrafine particles and elemental carbon.

CONCLUSIONS
Our observations serve as a demonstration and explanation of the epidemiologic evidence that associates the degree of traffic exposure with lung function in asthma.
AIR POLLUTION FROM ROAD TRAFFIC IS a serious health hazard, and particulates from diesel exhaust have become cause for increasing concern. Epidemiologic studies have demonstrated associations between ambient particulate matter and respiratory-associated morbidity and mortality; these effects may be greater among persons with preexisting respiratory disease, including asthma.1,2

Diesel engines emit relatively low concentrations of carbon monoxide and carbon dioxide, but as compared with gasoline engines of similar size, diesel engines can generate more than 100 times the number of particles per distance traveled3 and are major contributors to atmospheric particulate pollution. In urban environments, almost 90% of traffic-generated particulate matter is from diesel exhaust.4,5

Many urban residents, including those with increased susceptibility to the effects of air pollution, have short-term exposure to diesel traffic during normal activities. Studies of humans in exposure chambers have shown that controlled exposure to diesel exhaust can provoke increased airway resistance6,7 and bronchial inflammatory changes.8-11 Such studies are limited by their artificial nature, however, and the small numbers of study participants have typically been healthy.

We explored the effects of roadside traffic exposure on people with mild or moderate asthma, of differing severity, on a busy city street where traffic is entirely diesel-powered. We tested the hypothesis that short-term, ambient exposures to diesel traffic would lead to a reduction in lung function and a worsening of symptoms, accompanied by increased inflammation in the lungs.

METHODS

PARTICIPANTS
Using advertising and volunteer databases, we recruited 60 adults with asthma, none of whom were smokers; 31 had mild asthma and 29 had moderate asthma, as defined by the Global Initiative for Asthma.12 Each participant had intermittent wheezing and airway hyperresponsiveness to methacholine; the concentration of methacholine required to provoke a 20% decrease (PC_{20}) in the forced expiratory volume in 1 second (FEV_{1}) was <8 mg per milliliter. The participants were studied during periods of asthmatic stability, defined by the absence of exacerbations, respiratory infections, and treatment with oral corticosteroids for 4 weeks or more. During the study, the participants took their usual asthma medications.

The project was approved by the ethics committee at Brompton Hospital, London, and the institutional review board at the Robert Wood Johnson Medical School, New Brunswick, New Jersey. Written informed consent was provided by all participants.

STUDY DESIGN

In this randomized, crossover study, participants walked for 2 hours (10:30 a.m. to 12:30 p.m.) along the western end of Oxford Street, London’s busiest shopping street, where only diesel-powered buses and taxicabs are permitted, or through the traffic-free, western part of the nearby 142-hectare (about 350-acre) Hyde Park. Participants walked about 6 km during each exposure, at a steady pace on predefined paths, resting for 15 minutes every half hour. Exposure sessions, separated by more than 3 weeks, were confined to weekdays between November and March (2003 to 2005) to avoid pollen seasons; rainy days were also avoided. Equal numbers of participants were randomly assigned to each exposure sequence. The study had an estimated 90% power to detect a minimum difference in FEV_{1} response of 5.7% in each asthma-severity group between the two sites.

CLINICAL MEASUREMENTS

Participants measured their peak expiratory flow on a mini–Wright meter and recorded asthma symptoms 1 week before and after each exposure session; during the week before each exposure, they wore a diffusion tube to measure individual exposure to nitrogen dioxide throughout the week. At Royal Brompton Hospital, baseline measurements of FEV_{1}, forced vital capacity (FVC), and forced expiratory flow at 25 to 75% of vital capacity (FEF_{25-75}) were performed with the use of a spirometer (Vitalograph); the fraction of exhaled nitric oxide (FE_{Nox}) was recorded with the use of a chemiluminescence analyzer (Aerocrine). Exhaled breath condensate from tidal breathing was obtained with the use of a standardized breath-condensate collector (ECoScreen, Jaeger), and its pH was measured after de-aeration with argon.

Participants were driven for approximately 10 minutes to the exposure sites in a gasoline-powered car. On arrival and hourly during each session, we made further spirometric measurements;
on completion of each session, we asked participants to record any asthmatic symptoms. After returning to the hospital, we repeated spirometric and FeNO measurements regularly for 5 hours and measured responsiveness to methacholine; participants recorded any asthma symptoms over the ensuing 12 hours. The next morning, we repeated measurements of lung function and FeNO and collected samples of sputum induced by inhalation of a 3% sodium chloride solution from an ultrasonic nebulizer. Total cell counts were performed on a homogenized sputum sample with the use of 0.1% dithiothreitol, and differential cell counts were performed on 400 nonsquamous cells on cytospin slides. Supernatants were kept at −80°C for analysis of interleukin-8, myeloperoxidase, and eosinophil cationic protein with the use of commercially available immunoassay kits (IL-8 DuoSet, R&D Systems; and TiterZyme and UniCAP, Pharmacia Diagnostics AB; respectively).

EXPOSURE MEASUREMENTS
Throughout each exposure session, we measured number concentrations of ultrafine particles, using a real-time condensation particle counter (Model 3007, TSI) equipped with a ribbon laser light-scattering optical system (range, 0 to 100,000 particles per cubic centimeter; accuracy, ±20%). We collected fine particles smaller than 2.5 μm in aerodynamic diameter (PM_{2.5}) on quartz-fiber filters, using an air sampler (16 liters per minute). The filters were used first to determine PM_{2.5} mass concentration gravimetrically and were then analyzed for elemental carbon according to National Institute for Occupational Safety and Health guidelines (method 5040) (Sunset Laboratory). With a sampling pump, we collected nitrogen dioxide on Sep-Pak cartridges coated with potassium hydroxide and triethanolamine and subsequently analyzed the sample using ion chromatography.\(^\text{13}\) Temperature and relative humidity sensors, along with all air monitors, were located on a pushcart beside the participants.

STATISTICAL ANALYSIS
Descriptive summaries of exposure and health outcomes included means (±SD) for normally distributed variables and medians with ranges for other variables. Correlations between pollutants were examined.

Associations between exposure and health outcomes were examined through comparative analysis (Oxford Street vs. Hyde Park) and pollutant-specific, exposure-response analyses. A repeated-measures, mixed-effects linear regression model was constructed to estimate average values for health outcomes, with the use of an interaction term for the categorical variables of site and time to examine the effect of the exposure site on changes in lung-function or inflammatory biomarkers.\(^\text{14}\) A random effect for individual participants accounted for similarities across sessions for each person. A spatial-power covariance structure was used to model correlations between the unequally spaced repeated measurements within each session with a decay in the strength of correlation, depending on the time between measurements. With the use of the Akaike information criterion, this structure proved to model the data adequately relative to an unstructured correlation matrix. Additional covariates, including temperature and relative humidity and, where necessary, age, sex, body-mass index, and race or ethnic group were entered as covariates. Type 3 F tests of the site–time interaction produced P values representing the overall significance of the effect of site on changes in responses over time; contrasts of means were used to examine differences between sites in changes from baseline to particular time points. To study potential effect modifiers, we tested interactions between them and site–time interaction and conducted stratified analyses.

Because pollutant concentrations varied within and between sites, we performed pollutant-specific, exposure-response analyses using mixed models, as described above. In these models, we regressed the percent change from baseline in each of the lung-function and inflammatory-biomarker variables against pollutant concentrations averaged over each exposure session. Separate, single-pollutant models for PM_{2.5}, elemental carbon, ultrafine particles, and nitrogen dioxide were generated. Subsequently, we used two-pollutant models, in which two of the four pollutants were analyzed simultaneously.

RESULTS

As anticipated, participants with mild asthma had a higher baseline FEV\(_1\) than those with moderate asthma. They were also less likely to report limited exercise tolerance or symptoms provoked by exercise or traffic fumes (Table 1).

Median weekly nitrogen dioxide exposures were
Effects of Diesel Traffic Exposure in Asthma

Table 1. Baseline Characteristics of the Study Participants.*

<table>
<thead>
<tr>
<th>Characteristic</th>
<th>All Participants (N = 60)</th>
<th>Participants with Mild Asthma (N = 31)</th>
<th>Participants with Moderate Asthma (N = 29)</th>
<th>P Value</th>
</tr>
</thead>
<tbody>
<tr>
<td>Female sex — no. (%)</td>
<td>29 (48)</td>
<td>14 (45)</td>
<td>15 (52)</td>
<td>0.61</td>
</tr>
<tr>
<td>Age — yr</td>
<td></td>
<td></td>
<td></td>
<td>0.13</td>
</tr>
<tr>
<td>Mean</td>
<td>32</td>
<td>31</td>
<td>34</td>
<td></td>
</tr>
<tr>
<td>Range</td>
<td>19–55</td>
<td>20–49</td>
<td>19–55</td>
<td></td>
</tr>
<tr>
<td>Height — cm</td>
<td>172±8.8</td>
<td>172±8.4</td>
<td>171±9.3</td>
<td>0.67</td>
</tr>
<tr>
<td>Body-mass index†</td>
<td>23.2±3.7</td>
<td>23.2±3.6</td>
<td>23.2±3.9</td>
<td>0.98</td>
</tr>
<tr>
<td>White race — no. (%)‡</td>
<td>47 (78)</td>
<td>26 (84)</td>
<td>21 (72)</td>
<td>0.28</td>
</tr>
<tr>
<td>FEV₁ — % of predicted value</td>
<td>88.9±10.8</td>
<td>93.4±6.9</td>
<td>84.1±12.3</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>Atopy — no. (%)§</td>
<td>42 (84)</td>
<td>24 (89)</td>
<td>18 (78)</td>
<td>0.31</td>
</tr>
<tr>
<td>Methacholine PC_{20} — mg/ml¶</td>
<td>2.82±2.47</td>
<td>2.73±2.43</td>
<td>2.92±2.56</td>
<td>0.78</td>
</tr>
<tr>
<td>Treatment with inhaled corticosteroids — no. (%)</td>
<td>37 (62)</td>
<td>12 (39)</td>
<td>25 (86)</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>Unlimited exercise tolerance — no. (%)</td>
<td>51 (85)</td>
<td>28 (90)</td>
<td>23 (79)</td>
<td>0.23</td>
</tr>
<tr>
<td>Asthma affected by exercise — no. (%)</td>
<td></td>
<td></td>
<td></td>
<td>0.27</td>
</tr>
<tr>
<td>Yes</td>
<td>44 (73)</td>
<td>20 (65)</td>
<td>24 (83)</td>
<td></td>
</tr>
<tr>
<td>Not sure</td>
<td>4 (7)</td>
<td>3 (10)</td>
<td>1 (3)</td>
<td></td>
</tr>
<tr>
<td>Asthma affected by traffic fumes — no. (%)</td>
<td></td>
<td></td>
<td></td>
<td>0.19</td>
</tr>
<tr>
<td>Yes</td>
<td>17 (28)</td>
<td>7 (23)</td>
<td>10 (34)</td>
<td></td>
</tr>
<tr>
<td>Not sure</td>
<td>30 (50)</td>
<td>19 (61)</td>
<td>11 (38)</td>
<td></td>
</tr>
</tbody>
</table>

* Plus–minus values are means ±SD. P values are for comparisons according to the severity of asthma. FEV₁ denotes forced expiratory volume in 1 second.
† The body-mass index is the weight in kilograms divided by the square of the height in meters.
‡ Race was self-reported.
§ Ten participants (six with moderate asthma) did not have skin-prick tests.
¶ Methacholine PC_{20} denotes the concentration of methacholine required to provoke a 20% decrease in the FEV₁.

not significantly different before exposure at the two sites (Table 2). Participants had significantly higher exposures to PM$_{2.5}$, ultrafine particles, elemental carbon, and nitrogen dioxide on Oxford Street than in Hyde Park. Spearman’s rank-correlation coefficients for exposures to the four pollutants were as follows: 0.58 for the correlation of nitrogen dioxide with both ultrafine particles and elemental carbon, 0.62 for the correlation of PM$_{2.5}$ with ultrafine particles, and 0.84 for the correlation of ultrafine particles with elemental carbon. (Complete information is provided in Table A1 of the Supplementary Appendix, available with the full text of this article at www.nejm.org.)

Average daily peak expiratory flow, symptoms, and use of asthma medication did not differ significantly in the week before exposures at the two sites. Similarly, there were no significant differences between the sites in mean FEV₁, FVC, or FEF$_{25–75}$ at baseline in the hospital or on arrival at the site (Table 2). On the whole, changes over time in mean FEV₁ and FVC differed significantly between the sites (P=0.04 and P=0.01, respectively), but there were no significant differences between sites in the changes in FEF$_{25–75}$. Mean FEV₁ (percent of the predicted value) was lower after 1 hour of exposure at both sites. The subsequent decrement was greater and more sustained for Oxford Street, with a maximal decline at 2 hours (−6.1%, vs. −1.9% for Hyde Park; P<0.001). The differences between sites remained significant at every time point (P<0.05) until 22 hours after exposure. Among participants with moderate asthma, the decline in FEV₁ was greater for Oxford Street than for Hyde Park, but the difference was not significant.

The pattern for mean FVC (percent of predicted value) was similar (Fig. 1), with a maximum drop after 2 hours (−5.4% for Oxford Street vs. −1.6% for Hyde Park, P<0.005). The differences in changes in FVC between the two sites were significant at each time point between 2 and 5 hours after
the start of exposure (P<0.05) and were significantly greater among participants with moderate asthma than among those with mild asthma (P=0.008). There were no significant differences in the mean change in $\text{FEF}_{25\text{-}75}$ (percent of predicted value) between sites or in the methacholine $\text{PC}_{20}$ value 5 hours after exposure.

Changes in respiratory symptoms during and after exposures were small. Participants with mild asthma reported more symptoms after exposures on Oxford Street (both immediately afterward and 5 hours afterward) than after exposures in Hyde Park, but the differences were not significant. The use of treatments for asthma relief over the 7-day period after exposure did not differ significantly between the two sites.

There were no significant differences in the changes in mean $\text{FeNO}$ between exposures on Ox-

<p>| Table 2. Exposure Measurements for Oxford Street and Hyde Park and Lung Function before Exposure. |</p>
<table>
<thead>
<tr>
<th>Variable</th>
<th>Oxford Street</th>
<th>Hyde Park</th>
<th>P Value</th>
</tr>
</thead>
<tbody>
<tr>
<td>Exposure</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Nitrogen dioxide in previous week (μg/m$^3$)</td>
<td>23.5</td>
<td>22.3</td>
<td>0.90</td>
</tr>
<tr>
<td>Temperature (°C)</td>
<td>10.8</td>
<td>9.1</td>
<td>0.04</td>
</tr>
<tr>
<td>Relative humidity (%)</td>
<td>66</td>
<td>76</td>
<td>0.03</td>
</tr>
<tr>
<td>PM$_{2.5}$ (μg/m$^3$)</td>
<td>28.3</td>
<td>11.9</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>Ultrafine particles (thousands/cm$^3$)</td>
<td>63.7</td>
<td>18.3</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>Elemental carbon (μg/m$^3$)</td>
<td>7.5</td>
<td>1.3</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>Nitrogen dioxide (μg/m$^3$)</td>
<td>142</td>
<td>21.7</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>PM$_{10}$ (μg/m$^3$)</td>
<td>125</td>
<td>72</td>
<td>0.03</td>
</tr>
<tr>
<td>Baseline lung function</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>$\text{FEV}_1$</td>
<td>93.8±11.0</td>
<td>92.2±11.4</td>
<td>0.44</td>
</tr>
<tr>
<td>FVC</td>
<td>103.5±12.4</td>
<td>102.8±11.8</td>
<td>0.76</td>
</tr>
<tr>
<td>$\text{FEF}_{25\text{-}75}$</td>
<td>65.6±16.8</td>
<td>63.4±18.6</td>
<td>0.51</td>
</tr>
</tbody>
</table>

* Plus–minus values are means ±SD. P values are for within-participant comparisons between Oxford Street and Hyde Park. PM$_{2.5}$ denotes particles less than 2.5 μm in aerodynamic diameter, PM$_{10}$ particles less than 10 μm in diameter, $\text{FEV}_1$ forced expiratory volume in 1 second, FVC forced vital capacity, and $\text{FEF}_{25\text{-}75}$ forced expiratory flow at 25 to 75% of vital capacity.

† Values are for fixed-site measurements only.
ford Street and those in Hyde Park. There were greater decreases in the pH of exhaled breath condensate after exposure on Oxford Street — 1.16%, as compared with 0.88% in Hyde Park at 3 hours, and −1.90% as compared with 0.04% in Hyde Park at 6 hours (Fig. 2). The changes were not significantly greater among participants with moderate asthma.

The sputum myeloperoxidase concentration at 24 hours was higher after exposure on Oxford Street (24.5 ng per milliliter) than after exposure in Hyde Park (4.2 ng per milliliter, P=0.01) (Fig. 2). Sputum neutrophil counts and interleukin-8 concentrations were strongly correlated with myeloperoxidase (P<0.001 for both comparisons), and each was higher after Oxford Street exposures; only in the case of sputum neutrophil counts (for all participants), however, was this difference significant. There were no significant differences in sputum eosinophil counts or eosinophil cationic...
Figure 2. Mean Changes in Exhaled Breath Condensate pH, Sputum Neutrophil Counts, Supernatant Interleukin-8, and Supernatant Myeloperoxidase before and after Exposure on Oxford Street and in Hyde Park.

Changes in exhaled breath condensate pH are shown in Panel A for all participants, in Panel B for those with mild asthma, and in Panel C for those with moderate asthma. Changes in sputum neutrophil counts, supernatant interleukin-8, and supernatant myeloperoxidase are shown in Panels D, E, and F, respectively, for all subjects. With the use of mixed-effect models to test for overall changes in exhaled breath condensate pH over time, \( P = 0.01 \) for all participants, \( P = 0.27 \) for participants with mild asthma, and \( P = 0.02 \) for participants with moderate asthma. T bars represent 95% CI.
protein levels. We found no consistent evidence of an interaction between these health outcomes and the other variables listed in Table 1 (and Table A2 in the Supplementary Appendix).

Reductions in FEV₁, FVC, FEF₂₅⁻₇₅, and exhaled breath condensate pH were associated with ultrafine-particle exposures at most time points (Fig. 3); these reductions were also associated, although less consistently, with elemental carbon exposures (for ultrafine particles, $P = 0.02$, $P = 0.27$, $P = 0.11$, and $P = 0.03$, respectively; for elemental carbon, $P = 0.04$, $P = 0.11$, $P = 0.32$, and $P = 0.24$, respectively). Elemental carbon exposures were also associated with increases in FEV₁ concentrations ($P = 0.06$). For nitrogen dioxide, the associations were similar but less pronounced; for PM$_{2.5}$, there were no consistent associations. Increased sputum myeloperoxidase concentrations were associated with ultrafine-particle exposure ($P = 0.03$). There were no associations between sputum cell counts or interleukin-8 and any of the pollutant constituents. After adjustment for copollutants in the two-pollutant models, the effects of ultrafine particles and elemental carbon remained the most consistent and were significant (Table A3 in the Supplementary Appendix).

**DISCUSSION**

We examined the functional and inflammatory effects on asthma of real-life exposure to roadside diesel traffic. In adults with asthma, walking for 2 hours at a leisurely pace along a street where only diesel-powered vehicles were permitted resulted in a significant but essentially asymptomatic reduction in lung function. Although the changes were small, they were greater than those provoked by walking in a nearby park and were more pronounced among study participants whose asthma was more severe. These changes were accompanied by inflammatory changes in sputum and exhaled breath condensate.

The inflammatory response was predominantly neutrophilic, with raised levels of myeloperoxidase and interleukin-8 in sputum supernatants; we did not detect, even in participants with more severe asthma, an eosinophilic response characteristic of asthmatic inflammation. In association with the drop in FEV₁, we found a twofold increase in exhaled-breath-condensate hydrogen ions after exposure on Oxford Street. Acute asthma may be accompanied by airway acidification, with an increase in hydrogen ions by a factor of more than 100, perhaps reflecting inhibition of local epithelial proton pumps during airway inflammation. Thus, intracellular acidosis related to the inflammatory process may be reflected in reduced airway pH.

It has been suggested that any harmful respiratory effects of acute exposure to diesel exhaust are attributable more to its particulate content than to its gaseous content. In addition to coarse particles (2.5 to 10 μm in diameter), diesel exhaust contains, in far greater numbers, ultrafine particles (<0.1 μm in diameter). In our study, the most consistent relationships between changes in respiratory variables and specific pollutant concentrations were for ultrafine particles and elemental carbon, a finding consistent with growing evidence that the adverse respiratory effects of diesel-generated particles are attributable to those in the very small size range. With their higher ratio of surface area to mass, ultrafine particles can adsorb greater fractions of potentially toxic substances onto their surface, and they are deposited more deeply and in greater numbers within the lung than are larger particles. Furthermore, the carbon core of elemental carbon particles is highly adsorptive. Differences in the concentrations of ultrafine particles and elemental carbon between Oxford Street and Hyde Park were substantially larger than the differences in concentration between PM$_{2.5}$ and particles less than 10 μm in diameter. However, our findings cannot be taken as a demonstration of a causal association with ultrafine particles and elemental carbon, since these may simply be a sensitive proxy for the entirety of a roadside diesel-traffic exposure, which is composed not only of the complex diesel exhaust mixture but also of resuspended coarse, thoracic particles (small enough to enter the thorax) from road dust and engine or tire debris, which we did not measure.

Previous studies of the direct effects of diesel exhaust on asthma in humans have been conducted under laboratory conditions, with fresh diesel fumes, from which gaseous constituents may have been removed (“scrubbed”), or reconstituted diesel exhaust material delivered to subjects in an exposure chamber. The findings of these studies have not been entirely consistent, and none of them have demonstrated an effect on spirometric lung function, despite the use of much higher concentrations of diesel particles than
Figure 3. Point Estimates and 95% CI of the Percent Change in Health End Points per Incremental Change in Pollutant Components. Changes in the forced expiratory volume in 1 second (FEV\textsubscript{1}), forced vital capacity (FVC), forced expiratory flow between 25 and 75% of vital capacity (FEF\textsubscript{25–75}), and the fraction of exhaled nitric oxide (F\textsubscript{e}NO) are shown according to incremental changes in the levels of ultrafine particles (Panel A), fine particles (Panel B), elemental carbon (Panel C), and nitrogen dioxide (Panel D).
those we measured on Oxford Street. The discrepancies between the findings of other studies and our own may be due to the complexity of the natural pollutant mixture we studied, with interactions between particulates and other pollutants, or to the possibility that reconstituted or scrubbed diesel exhaust mixtures substantially reduce the proportion of ultrafine particles. In addition, we studied people with more severe asthma. Svartengren et al., in a realistic study of brief exposure to traffic pollution in a road tunnel, also failed to find any effect on FEV\textsubscript{1}, but reported more symptoms than we observed. Personal exposures to PM\textsubscript{2.5} and nitrogen dioxide were much higher in that study, but the study participants had milder asthma than did the participants in our study. On the other hand, the inflammatory responses we observed, particularly the increase in sputum neutrophils and myeloperoxidase, were similar in nature to those in studies of controlled, chamber exposures.\textsuperscript{6,10,26,28}

Our primary motive was to understand the effects of typical exposures to an urban atmosphere dominated by diesel exhaust. To this end, we selected a setting where only diesel traffic was allowed and used a randomized, crossover design limited to the winter months to avoid confounding exposure to pollens, thereby removing or controlling for most important confounding factors. Nonetheless, we recognize some limitations of our study, including the impossibility of blinding participants and the inability to exclude the possibility of subjective responses, particularly with symptom reporting. However, the internal consistency of our findings and the changes in variables over which the participants had no control make it very improbable that our results arose entirely from a subjective bias. Another difficulty arises from potentially confounding exposures that we did not measure. For example, a walk on Oxford Street is likely to be a more stressful experience than a walk in a quiet park, and it is possible that some of the responses we measured were induced by factors associated with stress, particularly noise.\textsuperscript{29} In addition, because we did not study a reference group of people without asthma, we cannot be sure that our findings are specific to people with asthma, although the more pronounced responses in the participants with more severe disease suggest that the findings are specific. Finally, we did not study exposure to gasoline-powered traffic and therefore cannot conclude that diesel traffic is more toxic than other types.

Our observations serve as a direct demonstration and explanation of the epidemiologic evidence that associates exposure to diesel traffic with the severity of asthma and of the symptoms that many patients with asthma report after exposure to diesel exhaust. The changes in our primary end point (FEV\textsubscript{1}) were small and unaccompanied by clinically significant symptoms but would be more important in patients with more compromised lung function. Without further study, however, we do not believe that these findings should deter most patients with asthma from visiting or working in busy urban environments. Our design has considerable advantages over orthodox chamber studies and could readily be adapted to assess therapeutic strategies in the prophylaxis of traffic responses in asthma or other cardiorespiratory diseases.

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Dr. Cullinan reports serving as a member of the Expert Advisory Panel for Air Quality Standards for the U.K. government. Dr. Chung reports receiving advisory-board fees from Astra-Zeneca, GlaxoSmithKline, Novartis, Merck, and Mundi-Pharma, consulting fees from Gilead, and research support from GlaxoSmithKline. No other potential conflicts of interest relevant to this article were reported.

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